Health Consultation

LINCOLN PARK COMPLEX DURRS NEIGHBORHOOD (OFF-SITE) SOIL

FT. LAUDERDALE, BROWARD COUNTY, FLORIDA

EPA FACILITY ID: FLN000407550

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U.S. DEPARTMENT OF HEALTH AND HUMAN SERVICES Public Health Service Agency for Toxic Substances and Disease Registry Division of Health Assessment and Consultation Atlanta, Georgia 30333

Health Consultation: A Note of Explanation

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In addition, consultations may recommend additional public health actions, such as conducting health surveillance activities to evaluate exposure or trends in adverse health outcomes; conducting biological indicators of exposure studies to assess exposure; and providing health education for health care providers and community members. This concludes the health consultation process for this site, unless additional information is obtained by ATSDR which, in the Agency's opinion, indicates a need to revise or append the conclusions previously issued.

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Prepared By:

Florida Department of Health Bureau of Community Environmental Health Under a Cooperative Agreement with the U.S. Department of Health and Human Services Agency for Toxic Substances and Disease Registry



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Summary

In this report, the Florida Department of Health reviewed 2004 and 2006 soil test results from the Durrs neighborhood at the request of the Florida Department of Environmental Protection. This neighborhood surrounds the 16.5-acre Lincoln Park Complex in Ft. Lauderdale, Florida. Previously, Florida DOH reviewed blood-lead test results and 2004 soil and groundwater test results.

The City of Ft. Lauderdale completely fenced the former incinerator area, cleaned contaminated soil on the "One-stop" property, and capped contaminated soil on the Lincoln Park playground. As a result, there are no current exposure pathways to soil contaminants on the Lincoln Park Complex site proper, as opposed to the potential for exposures to soil chemicals offsite in Durrs neighborhood yards and road right-of-ways, and other non-residential properties.

Chemicals found in off-site soil could have originated on the Lincoln Park Complex site as incinerator ash, wastewater treatment plant sludge, or from wastewater treatment plant flooding. Primary or secondary sources of chemicals measured in off-site soil could be residues from gasoline or diesel fuel combustion, asphalt roads or roofing materials, and residues from residential burning.

Although offsite soil testing is limited, the locations and levels of the chemicals measured seem to indicate incinerator ash may have been locally used as fill, because the highest arsenic and lead levels were measured in subsurface soil. In contrast, the highest dioxin and polychlorinated aromatic hydrocarbon (PAH) levels were measured in surface soil. The occurrence of the highest dioxin levels near the site and in surficial soil may indicate dioxins moved off the site as residues when the wastewater treatment plant flooded. PAH levels may not be related to the site; their levels tended to be only slightly elevated and were highest near roads in surface soil, additionally the highest PAH levels were not measured nearest the site.

The extent of soil contamination in the Durrs neighborhood needs to be adequately determined. Residents who dig (or have dug) in their yards may bring (or may have brought) **subsurface** incinerator residues to the surface where they are more likely to be contacted. Once the areas and depths of soil contamination are known, measures can be taken to prevent contact with soils having elevated chemical contaminant levels.

Although elevated chemical levels have not been measured in the surface soils of properties that are currently residential, persons walking in bare feet or sandals could contact contaminated **surface** soil on vacant properties or on rights-or ways near the Lincoln Park Complex that have elevated contamination, especially children that might then accidentally eat (ingest) the contacted soil. Depending on where soil contaminants are measured and at what levels, removal operations, deed restrictions, and/or engineering controls can be used to prevent future chemical exposures, for workers on City property and areas accessing utility lines, and for residents.

Until the full extent of soil contamination is determined, The Florida DOH recommends area residents and workers should avoid contact with soil that contains ash, glass, or metal pieces that might allow them to accidentally eat (ingest) soil and inhale dust. In addition as a precaution, residents should:

- use safe gardening practices (Appendix C); and
- only grow edible fruits and vegetables using raised beds with clean soil or compost.



Purpose

The Florida Department of Health (DOH) evaluates the public health significance of environmental contamination through a cooperative agreement with the federal Agency for Toxic Substances and Disease Registry (ATSDR) in Atlanta, Georgia. In March 2005, the Florida DOH reported the results of blood lead testing of children attending the former Lincoln Park School at the Lincoln Park Complex in Ft. Lauderdale. In October 2005, Florida DOH assessed the public health threat from soil and groundwater in the Durrs neighborhood surrounding the Lincoln Park Complex. At the request of the Florida Department of Environmental Protection (DEP), Florida DOH reviewed the additional 2006 soil test results together with the 2004 Durrs neighborhood soil testing results in this report.

Background

The 16.5-acre Lincoln Park Complex is in a commercial and light industrial area north of Sistrunk Boulevard between NW 18th Avenue and Interstate 95, in Ft. Lauderdale (Broward County), Florida (Figure 1). The complex includes the former City of Ft. Lauderdale (City) municipal incinerator and later a waste water treatment plant, the former Lincoln Park School (now the One-stop Shop for City of Ft. Lauderdale permits), and the Lincoln Park playground. Remediation of the playground and One-stop properties and the fencing of the incinerator propriety prevent exposure to the on-site soil contamination.

At a November 2003 public meeting, the Florida Department of Environmental Protection (DEP) shared the results of environmental testing in the Lincoln Park Complex with nearby residents. Florida DEP found layers of ash in soil deeper than one to two feet. Residents living north of the complex reported finding similar buried debris in their yards. In response, Florida DEP tested 30 surface soil (0-3") and 10 subsurface residential soil samples north of the complex in July 2004.

In a March 2005 health consultation report, Florida DOH found the blood lead levels in 40 children and three young adults attending or playing at the former Lincoln Park School were below the CDC guidelines for intervention for lead-poisoned individuals (ATSDR 2005d).

In an October 2005 public health assessment report, Florida DOH reviewed all of the 2004 soil and groundwater test data for the Lincoln Park Complex (ATSDR 2005a). We found the site posed "no apparent public health hazard," based on the information available at that time. We also found the health threat from past exposures was "indeterminate." In the public health assessment, Florida DOH addressed arsenic, copper, dioxins and polycyclic aromatic hydrocarbons (PAHs) in on-site surface soil (0-6"); and arsenic, copper, lead, and PAHs in on-site subsurface soil (3-24"). We concluded that recent exposures to surface soil are unlikely to have caused non-cancer illness. We also found "no apparent" increased theoretical cancer risk using the highest levels of all the chemicals measured in surface soil on the site.

In 2005, the City of Ft. Lauderdale secured grant funding for additional off-site soil testing. In February and March 2006, the City funded Florida DEP's contractor to test 25 more surface and subsurface soil samples in the surrounding Durrs neighborhood. In June 2006, the Florida DEP asked the Florida DOH to evaluate these soil test results. In this health consultation, we evaluate these results in combination with the 2004 soil test results.

In 2000, about 19,643 persons lived within a 1-mile radius of the site. Approximately 39% were 19 years of age or less. Approximately 88% were black, 8% were white, and less than 3% were



Latino/Hispanic. American Indian/Alaska Native, Asian/Pacific Islander, and all other racial/ethnic groups made up about 1% of the population (US Census Bureau 2000).

Community Health Concerns

Some residents in the Durrs neighborhood north of the Lincoln Park complex are concerned about the potential for contaminants in debris and waste materials they found buried in their yards.

Discussion

In this report, Florida DOH evaluates 2004 and 2006 test results from Durrs neighborhood soil. Florida DOH estimated soil ingestion and dust inhalation levels based on studies that measured people's actual exposure levels via soil dust and ingestion using a computer program (Risk Assistant 1.1). We are not aware of any other completed exposure pathways. Site-related contamination has not been traced to off-site surface water.

Although the City's contractor recently identified limited groundwater contamination in a monitoring well on the southeastern part of the site, municipal water is currently used by homes and businesses in the area, and drinking water wells are not likely to have been a past exposure pathway. However, some residents currently use irrigation wells. Although site-related contamination has not been measured in off-site groundwater, irrigation wells may be shallow and could pull in surface water. Because surface water may contain bacteria, DOH recommends people should not drink water from shallow irrigation wells or use it for bathing, cleaning food contact surfaces (like grills, dishes, or grilling utensils), or for rinsing food.

In July 2004, Florida DEP's contractor collected surface soil samples at 30 locations and subsurface soil samples at 10 of those locations. They analyzed all the surface samples for polycyclic aromatic hydrocarbons (PAHs), metals and dioxins. They analyzed all the subsurface samples for PAHs and metals and analyzed five samples for dioxins. They collected twenty-four **surface** soil samples on residential properties, half on City-owned, vacant or right-of-way properties which we designate "other" properties. They collected 4 **subsurface** soil samples on residential properties.

In February and March 2006, Florida DEP's contractor collected **surface** and **subsurface** soil samples at 25 locations. They analyzed all the surface samples for PAHs and metals, and 16 of the 25 samples for dioxins. They analyzed all the subsurface samples for PAHs and metals, and 15 of the 25 samples for dioxins. They collected nineteen **surface** and **subsurface** soil samples on residential properties, and 6 on "other" properties.

Levels of arsenic, dioxins, lead, and PAHs in some Durrs neighborhood soil samples were above ATSDR screening values. Arsenic, dioxins, lead and PAHs are chemicals that could be associated with incinerator ash. Use of chlorine in wastewater treatment can also produce dioxins. We discuss the public health threat of these contaminants below. The highest soil concentration of barium is below the ATSDR screening values and is thus not likely to cause illness.

We separated the soil data into yards currently being used as residences and "other" areas (vacant lots, City property, and rights-of-way). In general, residential yards had lower contaminant levels than "other" areas. Surface soil (0-3 inches deep) in residential yards had lower levels of arsenic, lead, and dioxins than "other" areas. Subsurface soil (3-24 inches deep)



in residential yards had lower levels of dioxins and lead than "other" areas. The highest level of arsenic, however, was in subsurface soils in a residential yard.

Figures 2 through 5 (Appendix A) show the soil sample locations where chemicals were measured above the ATSDR 1 in 1 million additional cancer risk screening guideline (CREGs)[†]. Florida DEP found arsenic, dioxins, and lead immediately north and east of the site with higher concentrations in subsurface than surface soil. Conversely, they found PAHs scattered throughout the area with higher concentrations in surface than subsurface soil. The distribution and depth of PAHs in soil suggest they may be associated with road asphalt, roofing tar or shingles, and vehicle emissions rather than incinerator ash. PAHs also tended to have leselevated levels, relative to their screening values.

Distribution of Arsenic and Dioxin in Durrs Neighborhood Surface/Subsurface Soil

Surface soil	Arsenic and dioxins >1 in 1 million cancer-risk soil value
(0-3")	38% (21/55)
Subsurface soil	Arsenic and dioxins >1 in 1 million cancer-risk soil value
(3-24")	49% (17/35)

Although contractors for Florida DEP have tested subsurface soil in 35 locations, the full extent of soil contamination in the Durrs neighborhood north and east of the site has still not been determined. Current data suggests that soil contamination may extend further to the north and east of the Lincoln Park Complex. Additional soil testing between NW 7th St and NW 7th Place, and one block east of NW 18th Place between NW6th Place and NW 7th Place (Figure 6) is necessary to assess the public health risk fully. Until the full extent of soil contamination is determined, residents living north and east of the Lincoln Park Complex (Figure 6) should avoid contact with soil that contains ash, glass, or metal pieces. Residents in this area that have ash, glass, or metal pieces in their soil should use the safe gardening practices outlined in Appendix C. Residents with ash, glass, or metal pieces in their soil should only grow edible fruits and vegetables using raised beds with clean soil or compost.

Public Health Implications

In this health consultation, we evaluate the risk of illness from exposure to soil in the surrounding Durrs neighborhood via incidental ingestion or inhalation of air-borne dust. We use a computer program (Risk Assistant 1.1) to estimate soil ingestion and dust inhalation levels.

Florida DOH evaluates chemical exposures by estimating daily doses for children and adults (Tables 5a, 5b, 6a, and 6b). A dose is an amount of chemical per body weight per day. Florida DOH compares estimated doses to amounts having known health effects from animal studies or from human medical reports (Tables 8 & 9). We use the units of milligrams (mg) of contaminant per kilogram (kg) of body weight per day (mg/kg/day). A milligram is 1/1,000 of a gram (a gram weighs about as much as a small raisin or paper clip); a kilogram is approximately 2 pounds.

 $[\]dagger$ **CREG**: ATSDR's Cancer Risk Evaluation Guide. A concentration in air, water, or soil (or other environmental media), which is derived from EPA's cancer slope factor and carcinogenic risk of 1x10-6 for oral exposure. It is the concentration that would be expected to cause no more than one excess cancer in a million persons exposed over a lifetime.



The following sections describe the relationships of the doses we calculated for the highest measured soil concentrations to health effects at the lowest doses known from human medical reports or animal studies. For each chemical, we considered:

- · Child and adult exposures
- Inhalation and ingestion exposures
- Residential and "other" location exposures
- Surface and subsurface soil exposures

Compared to ingestion, the estimated dose for inhalation was insignificant. This means that, if the ingestion and inhalation doses are added together, the inhalation doses insignificantly increase the entire exposure dose. Therefore, we only considered the ingestion doses for exposure. **Because we considered so many exposure scenarios (Tables 8 and 9), we compare only the highest estimated doses with the lowest doses known to cause illness in the following sections. Similarly, for cancer causing chemicals we discuss only the largest theoretical increased cancer risk.** We assumed residents are exposed to subsurface soil (3-24 inches deep) although routine contact with subsurface soil is unlikely.

Arsenic

Numerous medical studies document adverse health effects of long-term ingestion of arseniccontaminated water. We compare the highest estimated daily soil dose for Durrs neighborhood children and adults to the results of these studies. Recent animal studies indicate that mammals absorb between a quarter and a third of the arsenic that they ingest with their food, which may mean that soil ingestion results in a lower absorbed dose of arsenic than would result from arsenic-contaminated water (DEP 2005).

The highest estimated arsenic dose was for a child's daily ingestion of residential subsurface soil. This dose, 0.0007 mg/kg/day, is 1/3 the dose associated with cerebrovascular disease and interruption of the blood supply to the brain (resulting in brain damage). This dose is also about 1/3 the dose associated with adverse skin effects known form contaminated drinking water. While recent studies (Wasserman et al. 2004) have shown children's mental deficits beginning at this dose, daily exposure to subsurface soil is unlikely and all the other estimated exposure levels are lower. The highest surface soil dose (for a non-residential property) was about half the subsurface dose; for residential property, the highest measured level was 10 times lower than the highest-measured subsurface dose.

Based on limited off-site sampling, it appears that if incinerator ash was used as fill, it may generally still be buried; therefore the assumptions we made when calculating doses for daily, long-term (chronic) exposures probably **are not being met with subsurface soil**. Florida DOH's primary motivation for estimating the likelihood of adverse health effects for chronic exposures at the estimated dose levels was to demonstrate the need for determining where offsite contaminants occur so that appropriate measures are taken to prevent potential future exposures, including in areas with contaminated subsurface soil. Depending on where soil contaminants are measured and at what levels, removal operations, deed restrictions, and/or engineering controls can be used to prevent future chemical exposures, for workers on City property and for residents.

Although chronic exposures to subsurface soils are currently unlikely, the theoretical increased cancer risk for an adult's daily exposure to arsenic in residential subsurface soil is 5 in 100,000,



between "low" and "no apparent increased risk". Different cancers are associated with different arsenic doses. From lowest to highest dose, chronic arsenic exposures in people have been linked to lung cancer, basal and squamous cell skin cancers, liver cancer (haemangioendothelioma), urinary tract cancers (bladder, kidney, ureter and all urethral cancers), and intraepidermal cancers. Intraepidermal is the name for the early pre-invasive form of squamous cell skin cancer (ATSDR 2006).

Dioxin

Dioxins are a family of chemicals with similar structures and differing toxicities. The toxicities of each congener (chemical family member) are related to the most well studied member (2,3,7,8-dibenzo-p-dioxin) and they are added together to give a toxicity equivalent (TEQ) for the entire family. Dioxins can be produced when materials containing chlorine such as plastics are burned. Primary and secondary treatment of sewage and wastewater as well as the use of oxidation ponds also creates dioxins. Municipal garbage incineration, burning of yard trash and wastewater treatment may have contributed to the dioxin found in soil near the Lincoln Park Complex.

The highest dioxin TEQ dose was estimated for children via surface soil ingestion at one of the "other" properties. Because this dose was 40 times lower than the dose associated with altered developmental, social, and reproductive effects in a long-term study of rhesus monkeys fed dioxins in their food, and children's daily exposures to non-residential property soils are less likely to occur daily, non-cancer health effects from the measured dioxin levels are unlikely. DEP's contractors measured dioxins at lower levels in surface soil at residential properties and in subsurface soils, see Tables 8 and 9, Appendix B.

While daily long-term exposures might be unlikely to surface soil dioxins on "other" properties, the theoretical increased cancer risk for an adult's daily exposure is 2 in 100,000, or "no apparent increased risk". For specific information addressing dioxin and cancer links, see Tables 8 and 9.

Lead

For lead, estimated blood levels more accurately predict health effects than traditional dose estimates. We used a simple model (ATSDR 2006) to estimate blood lead levels from exposure to the highest levels of lead in Durrs neighborhood soil (Table 1). This model takes into account children and adults' exposures to lead from sources other than soil. Using this model, DOH assumed people might be exposed to lead-contaminated soil 8 hours per day. While this might be an exceptional assumption for other parts of the United States, it might be reasonable for persons not working outside the home, given the enjoyable climate and weather in Ft. Lauderdale. Table 7 (Appendix B) lists other model assumptions.



Table 1. Modeled (Estimated) Blood Lead Levels

	Child [†]	Adult [†]
Exposure	Blood Lead	Blood Lead
	(µg/dl)	(µg/dl)
Residential surface soil (maximum = 290 mg/kg)	1.5 – 3.3	0.8 – 2.9
"Other" surface soil (maximum = 330 mg/kg)	1.5 - 3.6	0.9 - 3.2
Residential subsurface soil (maximum = 870 mg/kg)	2.6-7.2	2.0-6.8
"Other" subsurface soil (maximum = 4,500 mg/kg)	9.8 - 31.1	9.2 - 30.7

mg/kg = milligrams per kilogram or parts per million (ppm)

 $\mu g/dl = micrograms per deciliter$

Current Centers for Disease Control Guidelines call for intervention when people's blood lead levels (BLLs) are above 10 μ g/dl. Intervention measures include determining where the person is contacting lead in their environments and abating that source. The BLLs FDOH modeled from the highest residential surface (290 ppm) and subsurface (870 ppm) soil lead levels are below this intervention BLL. Only the non-residential subsurface soil contained lead (4,500 ppm) at a level that might increase children's or adult's BLLs over this guideline; however, since this soil is non-residential and subsurface, **chronic exposures are currently unlikely.**

We discuss the potential for adverse health effects from chronic exposures to soil *the model shows <u>might be below CDC's intervention levels</u>, because:*

- the adverse health effects documented for low blood-lead levels include conditions that could exist for other reasons in the community, therefore chronic exposures to soil with lead might add to the causes of these conditions, off-site soil testing is incomplete, and
- we saw two gardens when we took residents' soil test results door-to-door.

Chronic exposures to the highest lead level measured in residential surface soil (290 mg/kg) might result in BLLs between 1.5 and 2.9 μ g/dl. In children, very low BLLs decreased the activity of an enzyme (ALAD) necessary for heme synthesis (no blood lead threshold level, Roels & Lauwerys 1987). Decreased heme synthesis can lead to anemia. Neurological and immunological effects occur in children at relatively low levels of exposure (1-17 μ g/dl, Altmann et al. 1998; and 1-45 μ g/dl Winneke et al. 1994). A BLL average of 2.3 μ g/dl increased blood pressure in adults (Den Hond et al. 2002).

Chronic exposures to the highest lead level measured in residential subsurface soil (870 mg/kg) may *currently* be much less likely than exposures to surface soil; however, if these soils were ever dug up, chronic exposures to them could have greater health impact because the lead level is higher. If excavated, chronic exposures to soil with the highest lead level might result in BLLs between 2.6 and 6.8 μ g/dl. In children, BLLs of:

- $5.4 \mu g/dl$ (Chiodo et al. 2001) were linked with decrements in attention, executive function, and visual-motor integration;
 - $7.7 \,\mu$ g/dl (Canfield et al. 2003) were linked with decline of 7.4 IQ points, and

[†] Most of the modeled levels are below Centers for Disease Control (CDC) guidelines of 10 ug/ dl for intervention when patients' actual blood lead levels are tested, adverse health effects are documented at blood lead levels below this intervention guideline level.



 $7-80 \mu g/dl$ (Angle and McIntire 1978; Angle et al. 1982) were linked with decreases in enzymes involved in cellular-level metabolism.

In adults, BLLs of:

- $3.3 \,\mu$ g/dl (mean, Muntner et al. 2003) were linked to decreased kidney function,
- $4.5 \,\mu\text{g/dl}$ (mean, Gennart et al. 1992) were linked to impaired cognitive performance (measurable, objective mental processes),
- \geq 5.1µg/dl (Torres-Sanchez et al. 1999) were linked to increased pre-term births to exposed mothers.

These modeled levels for residential subsurface soil do not represent levels people are likely to be exposed to currently, especially on a daily basis. Rather, they represent the potential for exposure, potential that may not be recognized by the community, and potential that may become more likely if residents dig on their property. For example, if residents plant shrubs and trees this might result in greater exposure potentials later, if contaminated buried materials are brought to the surface. The higher exposure levels that might come from currently buried materials adds urgency to the need to characterize offsite areas with incinerator ash and the need for implementation of mechanisms to prevent future exposures.

Currently, daily exposures to subsurface soil (the highest measured lead level was 4,500 ppm) on vacant properties, City-owned properties, or road right-of-ways, **probably would be less likely than the levels we modeled**^{Ω} and would probably be work-related. However, if these materials were ever excavated, chronic exposures could result in the highest modeled BLLs (9.8-30.7 µg/dl). As with residential subsurface soil, we modeled potential BLLs for chronic exposures to show that the areas with contaminated soils need to be identified (and addressed), to anticipate and prevent potential future exposures. In children, in addition to the results described above for lower BLLs, BLLs of:

- 11.9 μg/dl (Bhattacharya et al. 1993) were linked with postural disequilibrium (dizziness when standing);
- 12-120 μg/dl (Mahaffey et al. 1982, Rosen et al. 19890) were linked with decreased Vitamin D metabolism, and
- 12-17 μg/dl (Bornschein et al. 1989, McMichael et al. 1996, Moore et al. 1982, Ward et al. 1987, Wibberly et al. 1977) were linked reduced birth weight and/or reduced gestational age, and increased incidence of stillbirth and neonatal death.

In adults, BLLs of:

- $> 10 \,\mu$ g/dl (mean, Muntner et al. 2003) were linked to increased incidence of miscarriages and stillbirths, and
- $36 \mu g/dl$ (mean, Chia et al. 1992) were linked with postural disequilibrium (dizziness when standing).

Blood lead levels above 10 μ g/dl have not been confirmed for residents near the Lincoln Park Complex site. The Broward County Health Department offered blood-lead testing in April 2004 to anyone who was concerned they might have had exposure to lead-contaminated soil associated with the Lincoln Park Complex site. The County Health Department publicized this free testing (for children six years of age and younger, living in the 33311 zip code) through a press release to major and community media outlets. Broward County Health Department staff provided the testing at the Sunrise Health Center—Edgar P. Mills Multipurpose Center at 900

 $^{^{\}Omega}$ Our model assumed daily exposure, for eight hours a day, a scenario unlikely with subsurface soil, but deliberately done to show why future exposures to this buried soil should be prevented.



NW 31^{st} Avenue in Fort Lauderdale. Testing was available on Wednesday April 7, 2004 from 8:00 a.m. to 11:00 a.m. and Thursday, April 8, 2004 from 4:00 p.m. at 7:00 p.m. Approximately 50 persons were tested, including some adults, but none had blood lead level greater than or equal to 10 μ g/dl.

The Environmental Protections Agency (EPA) considers lead to be a probable human carcinogen. While worker studies have shown limited associations between elemental lead exposure and lung, stomach, kidney, and gliomal (brain and spinal cord) cancers in humans, a dose-response relationship has not been established and a cancer slope factor has not been calculated. Therefore, we were unable to calculate a lifetime excess cancer risks these estimated lead exposures.

Polycyclic Aromatic Hydrocarbons (PAHs)

Like dioxins, PAHs are a family of chlorinated compounds formed when organic chemicals (garbage, coal, oil, gasoline, wood, tobacco, and charbroiled meat) are burned. They are also found in asphalt, crude oil, coal, coal tar pitch, creosote, and roofing tar. To evaluate toxicity, we relate the toxicities of the carcinogenic PAH family members to the toxicity of benzo(a)pyrene, and then add them together for the PAH toxicity equivalent (TEQ).

The location of PAHs in surface soil (Figure 3) suggests they might not be exclusively site related. PAHs in soil may be more related to the proximity of asphalt and vehicle emissions from roads. Only 3 surface soil PAH TEQs were greater than 10 times the ATSDR 1 in 1 million increased cancer risk screening value. One soil sample appeared to contain roofing tar or asphalt. While 25 of 45 surface soil samples contained PAHs above the screening value, only 7 of 35 **subsurface** samples contained elevated PAHs. These 7 are located near roads rather than near the site (Figure 5).

The theoretical increased cancer risk for residential children's and adults' exposures to subsurface soil with the highest PAH TEQ level is 2 in 100,000, or "no apparent" increased risk, all other measured levels are lower and are listed in Tables 8 and 9.

Exposures to Mixtures

Dioxins (Schantz et al. 1992), arsenic (Wasserman et al. 2004) and lead (Chiodo et al. 2004) have been linked with developmental decrements at or near the lowest levels of exposure having reported health effects. Although the off-site sampling has been limited:

- 1 surface and 3 subsurface soil samples contained dioxin sufficient to produce children's doses above the minimum risk level calculated from a study showing adverse developmental effects in monkeys,
- 1 subsurface soil sample contained arsenic at a level that causes developmental and cognitive effects in children drinking arsenic-contaminated water, and
- 2 surface and 6 subsurface soil samples contained lead at levels that could affect children's and adults' cognitive processes.

Daily, long-term exposures to more than one of these chemicals in soil could have additive effects. Complete characterization of off-site soil contamination will assure that the future measures taken adequately safeguard children from possible developmental decrements.



Cancers from Mixtures - Although the highest individual chemical levels were not measured in samples from the same locations, we added the cancer risks for residential yards and "other" properties. While the theoretical increased cancer risk is higher for subsurface soil, people are not likely to have daily exposure to it. The theoretical increased cancer risk from exposure to the mixture to chemicals found in residential yards is between "no apparent increase" (1 in 100,000) and "low increased risk" (1 in 10,000). Arsenic, dioxin, lead, and PAHs have been linked to lung cancer. Arsenic dioxin and PAHs have been linked to skin cancer. In additions, arsenic has been linked to liver, gastrointestinal, and urinary tract cancers (stomach, intestines, kidney, and bladder). PAHs have been linked to bladder and other cancers. Lead has been linked to kidney, stomach, brain and spinal cord cancers. Dioxin has been linked with sort-tissue sarcoma, non-Hodgkin's lymphomas, liver and thyroid cancers.

Increased Theoretical Cancer Ri	isk from Chemical Mixture
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	Residential Yards	"Other" Properties
• Surface Soil	"no apparent increase", r 3 in 100,000 r	"low increased risk" 5 in 100,000
• Subsurface Soil	"low increased risk" $^{\Upsilon}$ 7 in 100,000 $^{\Upsilon}$	"low increased risk" 5 in 100,000

Quality Assurance and Quality Control

The completeness and reliability of the referenced environmental data determine the validity of the analyses and conclusions drawn for this health consultation. Florida DOH used existing environmental data. We assume these data are valid: Florida DEP's contractor and the laboratory they used have approved comprehensive quality assurance project plans.

Child Health Considerations

ATSDR and FDOH recognize the unique vulnerabilities of infants and children demand special attention (ATSDR 2005b). Children are at a greater risk than are adults for some hazardous substance exposures. Because children are smaller than adults are, their exposures can result in higher doses of chemical per body weight. If toxic exposures occur during critical growth stages, the developing body systems of children can sustain permanent damage. Probably most important, however, is that children depend on adults for risk identification and risk management, hygiene awareness, and access to medical care. Thus, adults should be aware of public health risks in their community so they can guide their children accordingly. In recognition of these concerns, ATSDR developed the chemical screening values for children's exposures that FDOH used in preparing this report to address site-specific child health considerations.

Susceptible populations have different or enhanced responses to toxic chemicals than most people. Reasons include genetic makeup, age, health, nutritional status, and exposure to other toxic substances (like cigarette smoke or alcohol). These factors may limit a susceptible person's ability to detoxify or excrete harmful chemicals or may increase the effects of damage to their organs or systems. For example, persons with the ALAD2 gene may be genetically susceptible to

^Y Total does not include the cancer risk from PAHs which may not be site related.

^r We did not use the cancer risk for the highest PAH value measured at one residence, because it was 2 in 10,000 and the other highest PAH values gave an increased risk of 2 in 1 million. FDOH asked FDEP about this PAH anomaly and they thought there might have been asphalt in this sample.



meningioma. Persons with the ALAD2 gene tend to have higher lead concentrations in their blood, and lead has been shown to increase the risk of brain cancer, particularly meningioma (Ranjaraman et al. 2005).

Conclusions

1. Based on the limited available data, the public health hazard category for surface soil contaminants on residential properties is "No Apparent Public Health Hazard" although daily exposures lasting for periods longer than a year (chronic exposures) to the highest levels of lead measured in surface soil might adversely affect sensitive subpopulations with high blood pressure and anemia. Arsenic, lead, and dioxin levels measured in residential surface soil might adversely affect sensitive subpopulations such as children with developmental disabilities.

Some of the levels of contaminants in residential surface and subsurface soil exceed the Florida residential Soil Target Cleanup Levels. The theoretical increased cancer risk from long-term exposures to a mixture of the highest measured levels of contaminants in residential surface soil is "low" to "no apparent."[†]

Arsenic and lead levels are generally higher in subsurface soils than in surface soils on residential properties, indicating incinerator ash may have been used as fill. Digging into subsurface soils on residential properties could increase residents' exposure potentials especially if buried wastes are brought to the surface.

2. Some off-site soil on non-residential properties that has been tested would be a "public health hazard" if people had daily, long-term exposures to it.

Modeling indicates daily exposures to off-site subsurface soils with the highest lead levels could result in blood lead levels above the Centers of Disease Control's action level (for intervention) of 10 μ g/L. While the model assumptions we made (daily, long-term exposures, lasting 8 hours a day) are unlikely to be met for buried soils that are not currently on residential property, these levels indicate **a potential** for excess lead exposure. The community may not recognize this potential, and workers who may replace or repair utility lines or cables buried in road right-of-ways, or may carry out construction on vacant properties, similarly may not recognize their potential for exposure.

3. Based on the distribution and the measured levels, polycyclic aromatic hydrocarbons (PAHs) in Durrs neighborhood soil do not appear to be related to the Lincoln Park Complex site.

Recommendations

1. Collect additional surface and subsurface soil samples from the Durrs neighborhood north and east of the Lincoln Park Complex (Figure 6). Analyze for arsenic, lead, and dioxins to characterize the extent and levels of contamination.

[†] Between 1 in 100,000 and 1 in 10,000 risk of an increased cancer case over the expected numbers of cases of all types of cancers combined.



- 2. Take measures to prevent future contamination exposures once the levels and extent of contamination are known.
- 3. Until the full extent of soil contamination is determined, avoid contact with soil in the untested areas north and east of the Lincoln Park Complex that contains ash, glass, or metal pieces (Figure 6). Residents in this untested area that have ash, glass, or metal pieces in their soil should use the safe gardening practices outlined in Appendix C. Residents with ash, glass, or metal pieces in their soil should only grow edible fruits and vegetables using raised beds with clean soil or compost. DEP has advised the residents whose yards have been tested and are known to have elevated levels of contaminants to follow these same recommendations.

Public Health Action Plan

Florida DOH will continue to work with DEP to inform and educate Durrs neighborhood residents.



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References

Altmann L, Sveinsson K, Kraemer U, et al. 1998. Visual functions in 6-year-old children in relation to lead and mercury levels. Neurotoxicology and Teratology 20(1):9-17.

Angle CR, McIntire MS. 1978. Low-level lead and inhibition of erythrocyte pyrimidine nucleotidase. Environ Res 17:296-302.

Angle CR, McIntire MS, Swanson MS, et al. 1982. Erythrocyte nucleotides in children increased blood lead and cytidine triphosphate. Pediatr Res 16:331-334.

[ATSDR] Agency for Toxic Substances and Disease Registry. 2006. Soil Comparison Values. Agency for Toxic Substances and Disease Registry, U.S. Department of Health and Human Services, Atlanta, GA.

[ATSDR] Agency for Toxic Substances and Disease Registry. 2006 Update. Toxicological profile for lead. Update. Atlanta: US Department of Health and Human Services.

[ATSDR] Agency for Toxic Substances and Disease Registry. 2005a. Public health assessment the Lincoln Park complex. Atlanta: U.S. Department of Health and Human Services.

[ATSDR] Agency for Toxic Substances and Disease Registry. 2005b. Public health assessment guidance manual. Atlanta: U.S. Department of Health and Human Services.

[ATSDR] Agency for Toxic Substances and Disease Registry. 2005d. Health Consultation, Lincoln Park Complex, Blood Lead Testing. Atlanta: U.S. Department of Health and Human Services.

[ATSDR] Agency for Toxic Substances and Disease Registry. 2004. Interaction profile for persistent chemicals found in fish. Atlanta: U.S. Department of Health and Human Services. February 2004.

[ATSDR] Agency for Toxic Substances and Disease Registry. 2002. Case study in environmental medicine. Lead toxicity. Atlanta: US Department of Health and Human Services.

[ATSDR]Agency for Toxic Substances and Disease Registry. 2000. Toxicological profile for arsenic. Update. Atlanta: U.S. Department of Health and Human Services. September 2000.

[ATSDR] Agency for Toxic Substances and Disease Registry. 1999. Toxicological profile for lead. Update. Atlanta: US Department of Health and Human Services.

[ATSDR] Agency for Toxic Substances and Disease Registry. 1998a. Toxicological profile for chlorinated dibenzo-p-dioxins. Update. Atlanta: U.S. Department of Health and Human Services.

[ATSDR] Agency for Toxic Substances and Disease Registry. 1998b. Guidance on Including Child Health Issues in Division of Health Assessment and Consultation Documents: U.S. Department of Health and Human Services. July 2, 1998

[ATSDR] Agency for Toxic Substances and Disease Registry. 1995. Toxicological profile for polycyclic aromatic hydrocarbons. Update. Atlanta: U.S. Department of Health and Human Services. Publication No.: PB/2000/108021



Baghurst PA, Robertson EF, McMichael AJ, et al. 1987. The Port Pirie cohort study: Lead effects on pregnancy outcome and early childhood development. Neurotoxicology 8:395-401.

Bhattacharya A, Smelser DT, Berger O, et al. 1998. The effect of succimer therapy in lead intoxication using postural balance as a measure: A case study in a nine-year-old child. Neurotoxicology (Little Rock) 19(1):57-64.

Bornschein RL, Grote J, Mitchell T, et al. 1989. Effects of prenatal lead exposure on infant size at birth. In: Smith M, Grant LD, Sors A, eds. Lead exposure and child development: An international assessment. Lancaster, UK: Kluwer Academic Publishers.

Bureau of the Census, 2000. LandView 5 Software on DVD, A Viewer for EPA, Census and USGS Data and Maps. U.S. Department of Commerce.

Canfield RL, Henderson CR, Cory-Slechta DA, et al. 2003. Intellectual impairment in children with blood lead concentrations below 10 micrograms per deciliter. N Engl J Med 348(16):1517-1526

Chia SE, Chia HP, Ong CN, Jeyaratnam J. 1996. Cumulative concentrations of blood lead and postural stability. Occup Environ Med 53(4):264-268.

Chiodo LM, Jacobson SW, Jacobson JL, 2004. Neurodevelopmental effects of postnatal lead exposure at very low levels. Neurotoxicol Teratol 26(3):359-371

Chiou H-Y, Huang W-I, Su C-L, et al. 1997. Dose-response relationship between prevalence of cerebrovascular disease and ingested inorganic arsenic. Stroke 28(9):1717-1723.

Den Hond E, Nawrot T, Staessen JA. 2002. The relationship between blood pressure and blood lead in NHANES III. J Hum Hypertens 16:563-568.

[DEP] Florida Department of Environmental Protection. 2005. Soil, Groundwater, and Surface Water Cleanup Target Levels (CTLs) for Chapter 62-777, Florida Administrative Code.

Ferreccio C, Gonzalez Psych C, Milosavjlevic Stat V, et al. 1998. Lung cancer and arsenic exposure in drinking water: a case-control study in northern Chile. Cad Saude Publica 14 (Suppl. 3):1993-198.

Gennert J-P, Buchet J-P, Roels H, et al. 1992. Fertility of male workers exposed to cadmium, lead or manganese. Am J Epidemiol 135:1208-1219.

Goodfellow F J L, V S G Murray, S K Ouki, A Iversn, A Sparks, T Bartlett. 2001. Public health response to an incident of secondary chemical contamination at a beach in the United Kingdom, Occupational Environmental Medicine, 2001; 58:232-238 (April)

Gupta P, Banerjee DK, Bhargava SK, et al. 1993. Prevalence of impaired lung function in rubber manufacturing factory workers exposed to benzo(a)pyrene and respirable particulate matter. Indoor Environ 2:26-31.

Hammond PB, Bornschein RL, Succop P. 1985. Dose-effect and dose-response relationships of blood lead to erythrocytic protoporphyrin in young children. In: Bornschein RL, Rabinowitz MB, Eds. The Second International Conference on Prospective Studies of Lead, Cincinnati, OH: April 1984. Environ Res 38:187-196.



Haque R, Mazumdar DN, Samanta S, et al. 2003. Arsenic in drinking water and skin lesions: Dose-response data from West Bengal, India. Epidemiology 14(2):174-182.

Hernberg S, Nikkanen J. 1970. Enzyme inhibition by lead under normal urban conditions. Lancet 1:63.

Hicks, Caldas LQA, Dare PRM, et al. 1986. Cardiotoxic and bronchoconstrictor effects of industrial metal fumes containing barium. Archives of Toxicology, Suppl 9. Toxic interfaces of neurons, smoke and genes, Secaucus, NJ: Springer-Verlag New York, Inc.

Hu H. 1991. Knowledge of diagnosis and reproductive history among survivors of childhood plumbism. Am J Public Health 81:1070-1072.

Ihrig MM, Shalat SL, Baynes C. 1998. A hospital-based case-control study of stillbirths and environmental exposure to arsenic using an atmospheric dispersion model linked to a geographical information system Epidemiology 9(3):290-294.

Lutz, PM, Wilson TJ, Ireland J, et al. 1999. Elevated immunoglobulin E (IgE) levels in children with exposure to environmental lead. Toxicology 134:63-78.

Lianfang W, Jianzhong H. 1994. Chronic arsenism from drinking water in some areas of Xinjiang, China. In: Nraigu JO ed. Arsenic in the environment: Part II Human health and ecosystem effects. New York, NY: John Wiley and Sons, Inc., 159-172.

Mahaffey KR, Rosen JF, Chesney RW, et al. 1982. Association between age, blood lead concentration, and serum 1,25-dihydroxycholecalciferol levels in children. Am J Clin Nutr 35:1327-1331.

McMichael AJ, Vimpani GV, Robertson EF, et al. 1986. The Port Pirie cohort study: Maternal blood lead and pregnancy outcome. J Epidemiol Community 40:18-25.

Moore MR, Goldberg A, Pocock SJ, et al. 1982. Some studies of maternal and infant lead exposure in Glasgow. Scott Med J 27:113-122.

Muntner P, He J, Vupputuri S, et al. 1983. Early appearance and localization of intranuclear inclusion in the segments of renal proximal tubules of rats following ingestion of lead. Br J Exp Pathol 64:144-155.

Neal J, Rigdon RH. 1967. Gastric tumors in mice fed benzo[a]pyrene: A quantitative study. Tex Rep Biol Med 25:553-557.

NTP 1994, Toxicology and carcinogenesis study of barium chloride dehydrate-(CAS no. 10326-27-0) in F344/N rats and B6C3F₁ mice, National Toxicology Program TR432

Nordstrom S, Beckman L, Nordensen I. 1979. Occupational and environmental risks in and around a smelter in northern Sweden: V. Spontaneous abortion among female employees and decreased birth weight in their offspring. Hereditas 90:291-296.

Perry HM Jr, Kopp SJ, Erlanger MW, et al. 1983. Cardiovascular effects of chronic barium ingestion. Trace Subst Environ Health 17:155-164.

Perry HM Jr, Perry EF, Erlanger MW, et al. 1985. Barium-induced hypertension. Adv Mod Environ Toxicol, Inorg Drinking Water Cardio Vasc Dis 9:221-229



Piomelli S, Seaman C, Zullow D, et al. 1982. Threshold for lead damage to heme synthesis in urban children. Proc Natl Acad Sci. 7:3335-3339.

Rabinowitz MB, Levilon A, Needleman H. 1986. Occurrence of elevated protoporphyrin levels in relation to lead burden in infants. Environ Res 39:253-257.

Rajaraman P, Schwartz BS, Rothman N, Yaeger M, Fine HA, Shapiro WR, Seiker RG, Black PM, and Inskip PD. 2005. δ-Aminolevulinic Acid Dehydratase Polymorphism and Risk of Brain Tumors in Adults. Environmental Health Perspectives. Sept 2005.

Rier SE, Martin DC, Bowman RE, et al. 1993. Endometriosis in Rhesus monkeys (Macaca mulatta) following chronic exposure to 2,3,7,8-tetrachlorodibenzo-p-dioxin. Fund Appl Toxicol 21:433-441.

Roels HA, Buchet J-P, Lauwerys R, et al. 1976. Impact of air pollution by lead on the heme biosynthetic pathway in school-age children. Arch Environ Health 31:310-316.

Roels HA, Lauwerys R. 1987. Evaluation of dose-effect and dose-response relationships for lead exposure in different Belgian population groups (fetus, child, adult men and women). Trace Elements in Medicine 4:80-87.

Rosen JF, Chesney RW, Hamstra AJ. et al. 1980. Reduction in 1,25-dihydroxyvitamin D in children with increased lead absorption. N Engl J Med 302:1128-1131.

Sanin LH, Gonzalez-Cossio T, Romieu I, et al. 2001. Effect of maternal lead burden on infant weight gain at one month of age among breastfed infants. Neurotoxicol Teratol 23:203-212.

Schantz SL, Ferguson SA, Bowman RE. 1992. Effects of 2,3,7,8-tetrachlorodibenzo-p-dioxin on behavior of monkey in peer groups. Neurotoxicol Teratol 14:433-446.

Torres-Sanchez LE, Berkowitz G, Lopez-Carrillo L, et al. 1999. Intrauterine lead exposure and preterm birth. Environ Res 81:297-301.

Tseng WP, Chu HM, How SW, et al. 1968. Prevalence of skin cancer in an endemic area of chronic arsenicism in Taiwan, J Nat'l Cancer Inst 40:453-463

Van Gennip AH, 1999. Defects in metabolism of purines and pyrimidines Ned Tijdschr Klin Chem 1999; 24: 171-175

Ward NI, Watson R, Bruce-Smith D. 1987. Placental element levels in relation to fetal development for obstetrically normal births: A study of 37 elements: Evidence for the effects of cadmium, lead, and zinc on fetal growth and for smoking as a source of cadmium. Int J Biosoc Res 9:63-81.

Wasserman, GA, X Liu, F Parves, H Ahsan, P Factor-Litvak, A van Geen, V Slavkovich, NJ LoIacono, Z Cheng, I Hussain, H Momataj, and JH Graziano. September 2004. Water Arsenic Exposure and Children's Intellectual Function in Araihazar, Bangladesh. Environmental Health Perspectives 112(13)1329-1333.

Wibberly DG, Khera AK, Edwards JH. et al. 1977. Lead levels in human placentae from normal and malformed births. J Med Genet 14:339-345.



Winneke G, Altmann L, Kramer U, et al. 1994. Neurobehavioral and neurophysiological observations in six-year-old children with low lead levels in East and West Germany. Neurotoxicology 15(3):705-713.

Wones RG, Stadler BL, Frohman LA. 1990. Lack of effect of drinking water barium on cardiovascular risk factors. Environ Health Perspect 85:355-359.

Appendix A—Figures













Appendix B—Tables



Table 2a. Completed exposure pathways

Pathway Name	Exposure Pathway Elements						
	Source	Environmental/ Exposure Media	Point of Exposure	Route of Exposure	Exposed Population and land use	Time	
Contaminated off-site surface soil, dust	Residential soil and soil on other properties	Wastes, surface and subsurface soil	Off-site properties	Incidental ingestion and inhalation	Off-site residents/owners, workers	Past Current Future	

Table 2b. Potential exposure pathways

			Exposure Pathway Elen	nents		
Pathway Name	Source	Environmental/ Exposure Media	Point of Exposure	Route of Exposure	Exposed Population and land use	Time
Contaminated off-site subsurface soil, dust	Residential soil and soil on other properties	Wastes, surface and subsurface soil	Off-site properties	Incidental ingestion and inhalation	Off-site residents/owners, workers	Past Current Future

Table 2c. Incomplete exposure pathways

Pathway Name	Exposure Pathway Elements						
	Source	Environmental/ Exposure Media	Point of Exposure	Route of Exposure	Exposed Population and land use	Time	
Shallow groundwater	Contaminated groundwater on the site	Shallow groundwater	Off-site irrigation well	Incidental ingestion and inhalation	Down-gradient residents	Future	



Contaminants	Screening Value (mg/kg) ATSDR:	DEP:	Highest Soil Concentration (mg/kg)		Location of Highest Concentration		Number Soil Samples Above Screening Value	
of Concern	Children/adults		residences	Other*	residences	Other*	residences	Other*
arsenic	20/200 EMEG	2.1 SCTL	6.3	30	Front yard of 2 nd res. N of former Inc. RSLPRSS-16	N boundary of former Inc. site LPRSS-9	ATSDR DEP 0/43, 12/43	ATSDR DEP 1/12, 6/12
barium	30,000 400,000 EMEG	210 SCTL	230	BSL/120	Back yard, lot N. of former inc. LPRSS-8	Right-of-way E. of former Incin. LPRSS-3	ATSDR DEP 0/43, 1/43	ATSDR DEP 0/12, 0/12
dioxin TEQ	0.00005/0.0007 EMEG 0.0	00007 SCTL	0.00003	0.0002	Side yard of 2 nd res. N of former Inc. RSLPRSS-19	Vacant lost N. of former Incin. LPRSS-10	ATSDR DEP 0/35, 6/35	atsdr dep 1/10, 5/10
lead		400 SCTL	BSL/290	BSL/330	Side yard of 2 nd res. N of former Inc. RSLPRSS-19	Right-of-way E. of former Incin. LPRSS-3	0/43	0/12
PAHs TEQ	0.1 CREG 0.1 resid	dential SCTL	51/3	0.447	Front yard of 2 nd res. N of former LP school RS2LPSS-8	Vacant lot N. of former Incin. RS2LPSS-21	21/43	4/12

Table 3. Maximum concentrations in off-site surface soil (0 to 3 inches below ground surface).

CREG—ATSDR's Cancer Risk Evaluation Guide for 1 excess cancer case in 1 million people (ATSDR 1992a).

EMEG—Environmental Media Evaluation Guide

Inc. or Incin.—Incinerator Property

mg/kg—milligrams per kilogram

N—north

PAHs—polycyclic aromatic hydrocarbons SCTL—FDEP's Soil Target Cleanup Level for residential land uses. * Other—sites include vacant lots, lots owned by the city, and road right-of-ways



Table 4. Maximum c	oncentrations in	off-site	subsurface	soil (3 to 2	24 inches	below ground surface).

Contaminants	Screening Value (mg/kg) ATSDR:	DEP:	Highest Soil Concentration (mg/kg)		Location of Highest Concentration		Number Soil Samples Above Screening Value	
or Concern	Children/adults		residences	Other*	residences	Other*	residences	Other*
arsenic	20/200 EMEG	2.1 SCTL	53	40	Side yard of 2 nd res. N of former Inc. RSLPRSS-19	Right-of-way E. of former Incin. LPRSSB-3	ATSDR DEP 1/23 8/23	ATSDR DEP 0/12 7/12
barium	30,000 400,000 EMEG	120 SCTL	520	2300	Side yard of 2 nd res. N of former Inc. RSLPRSS-19	Right-of-way E. of former Incin. LPRSSB-3	ATSDR DEP 0/23 3/23	ATSDR DEP 0/12 2/12
dioxin TEQ	0.00005/0.0007 EMEG 0.0	00007 SCTL	0.00008	0.0001	RS2LPSB-13	Vacant lot N. of former Incin. LPRSS-10	ATSDR DEP 2/14 6/14	atsdr dep 2/7 4/7
lead		400 SCTL	870	4,500	Side yard of 2 nd res. N of former Inc. RSLPRSS-19	Right-of-way E. of former Incin. LPRSSB-3	dep 3/23	DEP 3/12
PAHs TEQ	0.1 CREG 0.1 resid	lential SCTL	18.3/1.3	0.143	Front yard of 2 nd res. N of former LP school RS2LPSB-8	Vacant lot N. of former Incin. RS2LPSS-21	7/23	1/12

CREG—ATSDR's Cancer Risk Evaluation Guide for 1 excess cancer case in 1 million people (ATSDR 1992a).

EMEG—Environmental Media Evaluation Guide

Inc. or Incin.—Incinerator Property

mg/kg—milligrams per kilogram

N—north

PAHs—polycyclic aromatic hydrocarbons

SCTL—FDEP's Soil Target Cleanup Level for residential land uses. * Other—sites include vacant lots, lots owned by the city, and road right-of-ways



- -									
	Contaminant of	Oral	Soil/dust-Ingestion	Inh					

Table 5a.	Estimated	doses from	exposures	to residentia	surface	soil.
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Contaminant of Concern	Oral MRL	Soil/dust (mg/k	Soil/dust-Ingestion (mg/kg/day)		Soil/dust- Inhalation (mg/m ³)
(maximum concentration: mg/kg)	(mg/kg/day)	Child	Adult	(mg/m^3)	Child and Adult
arsenic (6.3)	0.0003 Chr	0.00008	0.000009	-	0.0000003
barium (230 mg/kg)	0.6 Chr	0.003	0.0003	-	0.00001
dioxin TEQ (0.00003)	0.000000001 Chr	0.0000000004	0.00000000004	-	0.00000000002
lead (290)	-	М	М	-	М
PAHs (51)/(3)	-	0.0007/0.00004	0.00007/0.0000002	-	0.000003/0.0000002

Table 5b. Estimated doses from exposures to right-of-way, city-owned or vacant lot surface soil.

Contaminant of Concern	Oral MRL	Oral Soil/dust- MRL (mg/k		Inhalation MRL	Soil/dust- Inhalation (mg/m ³)
(maximum concentration mg/kg)	(mg/kg/day)	Child	Adult	(mg/m^3)	Child and Adult
arsenic (30)	0.0003 Chr	0.0004	0.00004	-	0.000002
barium (120 mg/kg)	0.6 Chr	0.002	0.0002	-	0.000007
dioxin TEQ (0.0002)	0.000000001 Chr	0.000000003	0.000000003	-	0.00000000001
lead (330)	-	М	М	_	М
PAHs (0.447)	-	0.000006	0.0000006	-	0.00000002



Table 6a. Estimated doses from exposures t	o residential subsurface soil.
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Contaminant of Concern	Oral MRL	Soil/dust-Ingestion (mg/kg/day)		Inhalation MRL	Soil/dust- Inhalation (mg/m ³)
(maximum concentration mg/kg)	(mg/kg/day)	Child	Adult	(mg/m^3)	Child and Adult
arsenic (53)	0.0003 Chr	0.0007	0.00008	-	0.000003
barium (520 mg/kg)	0.6 Chr	0.007	0.0007	-	0.00003
dioxin TEQ (0.00008)	0.000000001 Chr	0.000000001	0.0000000001	-	0.000000000004
lead (870)	-	М	М	-	М
PAHs (18.3)(1.3)	-	0.0002/	0.00003/	-	0.000001/0.00000007

Table 6b. Estimated doses from exposures to right-of-way, city-owned or vacant lot subsurface soil.

Contaminant of Concern	Oral MRL	Oral Soil/dust MRL (mg/k		Inhalation MRL	Soil/dust- Inhalation (mg/m ³)
(maximum concentration mg/kg)	(mg/kg/day)	Child	Adult	(mg/m^3)	Child and Adult
arsenic (40)	0.0003 Chr	0.0005	0.00006	-	0.000002
barium (2300 mg/kg)	0.6 Chr	0.03	0.003	-	0.0001
dioxin TEQ (0.0001)	0.000000001 Chr	0.000000001	0.0000000001	-	0.000000000006
lead (4,500)	-	М	М	_	М
PAHs (0.143)	-	0.000002	0.0000002	-	0.00000008



Explanations for abbreviations and footnotes used on Tables 5 and 6.

- Acute Acute exposure length of 0-14 days
- Int Intermediate exposure length of 15- 364 days
- Chr Chronic exposure length of more than 365 days
- N.S. Not Significant
- mg/kg/day milligram chemical per kilogram body weight per day
- mg/m3 microgram of chemical per cubic meter of air
- M values were modeled (tables precede these explanations)

MRL – Minimum Risk Level extrapolation of a No Observable Adverse Effect level in a study to exposures, calculated by dividing the study dose by safety factors.



 Table 7. Estimated Blood Lead Concentrations In Children and Adults Ingesting

 Off-Site

(0 to 3" foot) Surface Soil and Subsurface Soil (3'-24") (micrograms per deciliter - µg/dl)

Media	Children	Adults
Off-site residential surface soil	1.5-3.3	0.8-2.9
Off-site other surface soil	1.5-3.6	0.9-3.2
Off-site residential subsurface soil	2.6-7.2	2.0-6.8
Off-site other subsurface soil	9.8-31.1	9.2-30.7

Values used to Estimated Blood Lead Concentrations in Persons Ingesting On-Site

Time 8 hrs a day for	Values for Slopes	children	Values for adults Slopes		
both	Low		Low	High	
0.33	2.46	0.33	1.59	3.56	
0.33	2.46	0.33	1.53	3.56	
0.33	0.24	0.33	0.016	0.0195	
0.33	0.16	0.33	0.03	0.06	
0.33	0.002	0.33	0.002	0.016	
0.33	0.004	0.33	0.004	0.004	

(0 to 2 foot) Soil (micrograms per deciliter - μ g/dl)

*Default Value from ATSDR 1999a, Appendix D.

These slopes were for children and adults from ATSDR 1999a, Appendix D.

ATSDR's Regression Analysis with Multiple-uptake Parameters to Estimate Blood Lead from Environmental Exposures (ATSDR 1999a, Appendix D)



Table 8. Comparison of doses calculated from highest measured surface soil values to lowest observable adverse effect levels (LOAELs) in animal and human medical studies. Bolded doses are above LOAEL or minimum risk level (MRL).

Chemical	Soil Doses are in mg/kg	g/day, inh. are in mg/m ³	theoretical incr	reased cancer risk
Chemical	children's dose	adults' dose	children's	adults'
Arsenic (residences)	Ing 0.00008 Inh 0.000003	Ing 0.000009 Inh 0.0000003	Ing 5:1,000,000 Inh <1:1,000,000	Ing 6:1,000,000 Inh <1:1,000,000
Arsenic (other [†])	Ing 0.0004 Inh 0.000002	Ing 0.00004 Inh 0.000002	Ing 3:100,000 Inh 1:1,000,000	Ing 3:100,000 Inh 2:1,000,000
ATSDR 2000 (Update)	Inn 0.00002 Child (residential) ingestion dose (associated with cerebrovascular didamaged brain tissue). Another stathan the MRL (0.0003) calculated term ingestion of arsenic in drinkin calculating the MRL. Adult (residential) ingestion dose in not expect skin or gastrointestinal Inhalation (residential) exposure let (0.0007, Ihrig et al., 1998, As 3 ⁺) arsenic. Dermatitis is skin inflamm Child (other) ingestion dose (0.000 cerebral infarction (an interruption the MRL (0.0003) calculated from drinking water. ATSDR scientists Sensitive children or children inge Adult (other) ingestion dose, 0.000 gastrointestinal health effects for a Inhalation (other) exposure level (0.0007, Ihrig et al., 1998, As 3 ⁺) arsenic. Dermatitis is skin inflamm Arsenic Associated cancers: From cancer, basal and squamous cell skin inflamm	(0.00008) is $1/25$ times the lowest of sease and cerebral infarction (an intr idy showed skin effects at 0.0018 (F from a no observable adverse effect ng water. ATSDR scientists divided is 8.8 times less than the (0.00008) of health effects for most adults and ch evel (0.0000003) is 2,333 times less and 23,333 times less than the dose of hation that may cause redness, pain, 04) is $1/5^{th}$ the LOAEL dose (0.002, in the blood supply to any part of the the NOAEL (0.0008, Tseng et al.19 divided this NOAEL dose (0.0008) esting more than the level of soil we 004, is 10 times less than the childred dults at this level of exposure. (0.000002) is 350 times less than the and 3,500 times less than the dose can hation that may cause redness, pain, lowest to highest dose cancer effect in cancers, liver cancer (haemangio	provide adverse effect (LOAEL) provide adverse effect (LOAEL) erruption of the blood supply to any Haque et al., 2003). This child resid level NOAEL (0.0008, Tseng et al this NOAEL dose (0.0008) by 3 to lose referenced for children and 1/2 hildren, at these exposure levels. than the amount associated with in causing dermatitis (0.007, Mohame and occasionally itching. Chiou et al. 1997) associated with in be brain, resulting in damaged brain 2068) for adverse skin effects from 1 by 3 to account for human diversity estimated might experience adverse en's (other) dose (0.0004), we woul and occasionally itching.	dose (0.002, Chiou et al. 1997) y part of the brain, resulting in ential ingestion dose is 3.75 less 1.1968) for skin effects from long- o account for human diversity in 222 nd the LOAEL dose, we would creased risk of still birth in humans ed 1998, As 3^+) in humans inhaling cerebrovascular disease and n tissue). This level is $1/_3$ more than ong-term ingestion of arsenic in y in calculating the MRL (0.0003). e skin effects. Id not expect skin or creased risk of still birth in humans 1 1998, As 3^+) in humans inhaling in people have been linked to lung rs (bladder, kidney, ureter, and all
right-of-ways, city- owned or vacant lot surface soil.	urethral cancers), and intraepiderm invasive means that the cancer cell have spread to the lymph nodes, by and spread into the lymphatic syste	that cancers, niver cancer (naemangio nal cancers. Intraepidermal is the nar ls are confined to the outermost laye ut they can spread along the skin sur em.	me for the early pre-invasive form or of skin, the epidermis. At this state fractional for the state of the s	of squamous cell skin cancer. Pre- ge, the cancer cells are unlikely to an develop into an invasive cancer



Chamical	Soil Doses are in mg/kg	g/day, inh. are in mg/m ³	theoretical increased cancer risk		
Chemical	children's dose	adults' dose	children's	adults'	
Barium (residences)	Ing 0.003 Inh 0.00001	Ing 0.0003 Inh 0.00001	No Slope No Unit Risk	No Slope No Unit Risk	
Barium (other [*])	Ing 0.002 Inh 0.000007	Ing 0.0002 Inh 0.000007	No Slope No Unit Risk	No Slope No Unit Risk	
ATSDR 2006 (Update) Draft *Other properties include right-of-ways, city- owned or vacant lot surface soil.	Child (residential) ingestion mg/kg/day, NTP 1994) with for this study was 65 mg/kg/ exposure by dividing the NC variability).Adult (residential) ingestion not expect kidney or other he Inhalation (other) exposure 1 of barium by inhalation. The exposed for an unspecified a bronchoconstriction. The est Child (other) ingestion dose mg/kg/day, NTP 1994) with for this study was 65 mg/kg/ exposure by dividing the NC variability).Adult (other) ingestion dose mg/kg/day, NTP 1994) with for this study was 65 mg/kg/ exposure by dividing the NC variability).Adult (other) ingestion dose expect kidney or other health Inhalation (other) exposure 1 of barium by inhalation. The exposed for an unspecified a bronchoconstriction. The est	dose (0.003) is 200 times less barium chloride in their food day. ATSDR authors used this DAEL by 100 (10 to account for dose (0.0003) is 10 times less ealth effects for most adults an <u>evel</u> Medical case reports and lowest reported exposure level mount of time to this concentri imated exposure level, (0.000 (0.002) is 300 times less than barium chloride in their food day. ATSDR authors used this DAEL by 100 (10 to account for (0.0002) is 10 times less than n effects for most adults and cl <u>evel</u> Medical case reports and lowest reported exposure level mount of time to this concentri imated exposure level, (0.000	than the chronic MRL (0.6 m for 90 days showed increased s NOAEL to derive the chron or extrapolation from animals than the (0.003) dose referen ad children, at these exposure animal studies are inadequate el is 0.06 mg/m ³ /minute (Hicl ration of aerosolized barium c 01) is 6,000 times less. the chronic MRL (0.6 mg/kg/ for 90 days showed increased s NOAEL to derive the chron or extrapolation from animals the (0.003) dose referenced ff nildren, at these exposure leve animal studies are inadequate el is 0.06 mg/m ³ /minute (Hicl ration of aerosolized barium c 007) is 8,571 times less.	ng/kg/day). Rats dosed (115 l kidney weight. The NOAEL ic MRL of 0.6 for barium to humans and 10 for human acced for children, we would levels. e to establish the heath effects ks et al. 1986). Guinea pigs chloride solution experienced /day). Rats dosed (115 l kidney weight. The NOAEL ic MRL of 0.6 for barium to humans and 10 for human for children, we would not els. e to establish the heath effects ks et al. 1986). Guinea pigs chloride solution experienced	
	Barium associated cancers. H	Barium has not been shown to	cause cancer in people or ani	mals.	



Chemical	Soil Doses are in mg/kg	Soil Doses are in mg/kg/day, inh. are in mg/m ³		theoretical increased cancer risk		
Chennear	children's dose	adults' dose	children's	adults'		
Dioxin TEQ (residences)	Ing 0.000000004 Inh 0.00000000002	Ing 0.00000000004 Inh 0.000000000002	Ing 3:1,000,000	Ing 3:1,000,000		
Dioxin TEQ (other [*])	Ing 0.00000003 Inh 0.00000000001	Ing 0.000000003 Inh 0.0000000001	Ing 2:100,000	Ing 2:100,000		
ATSDR 1998a (Update)	Child res. ingestion dose (0.0000 (moderate endometriosis) and alt of 0.000000001 calculated from neurobehavioral and developmen LOAEL, 3 for extrapolation from that the effects that occur at the I People's ingestion exposures are <u>Adult res. ingestion dose</u> (0.0000 most adults and children at these <u>Child other ingestion dose</u> (0.0000 endometriosis) and altered social While this level might indicate a daily. <u>Adult other ingestion dose</u> (0.0000 children, it is unlikely adults won <u>Inhalation</u> of dioxins has not bee inhalation and dermal exposure, levels of exposure are associated Higher levels are associated with diabetes. The highest exposure levels	0000004) is 1/300 th the dose (0.000 tered social behavior in a dioxin rh the 0.00000012 dose (Schantz et a ntal effects. ATSDR authors divide n animals to humans and 10 for hu owest levels of dioxin doses are in mainly known from low levels of 00000004) is 10 times less than the residential exposure levels. 0000003) is 1/40 th the dose (0.0000 l behavior in a dioxin rhesus monk n increased risk for children expose 00000003) is 400 times less than the ald experience health effects from n studied in animals. People's occ but health effects are known prima with hormone changes that can re a immunosuppression, changes in t evels are associated with nervous s	200012, Rier et al., 1993) associat resus monkey study. This dose is al., 1992) at which rhesus monkey ed this dose by an uncertainty fac- man variability. The results of other nmune and endocrine, in addition food contamination. e residential children's dose. We we 200012) associated with reproduct they study. However, this dose is 3 and daily, such exposures to the "of the (0.00000012) LOAEL health exposures at these levels. upational and accidental exposure arily from associations with the level sult in changes in sex ratios in chi- the liver, abnormal glucose tolera system effects, chloracne, respirat	ted with reproductive effects also 2.5 time less than the MRL vs showed adverse tor of 90, 3 for use of a minimal her oral animal studies suggest to developmental effects. would not expect health effects for ive effects (moderate times greater than the MRL. other" properties might not occur effects described above for es to dioxin involve primarily evels stored in fat. The lowest hildren (more females are born). nce, and increased risk of ory effects, and increased risk of		
*Other properties include right-of-ways, city- owned or vacant lot surface soil.	cancer. <u>Cancers</u> Statistically significant is periods. Although the estimated consistent across studies with the with some data suggesting a poss- including lung cancer. Animal st	increases in risks for all cancers we Standardized Mortality Ratios (a r highest dioxin exposures. The ev sible relationship between soft-tiss udies have also shown association	ere found in workers highly expo atio that is a direct comparison idence linking peoples' doses wit ue sarcoma, non-Hodgkin's lymp s with liver, thyroid and skin can	sed to dioxins with longer latency with a standard) are low, they are h site-specific cancers is weaker, whoma, or respiratory cancers cer.		



Chamical	Soil Doses are in mg/kg/day, inh. are in mg/m ³		theoretical increased cancer risk		
Chemical	children's dose	adults' dose	children's	adults'	
PAHs TEQ (residences)	Ing 0.0007/0.00004	Ing 0.00007/0.0000002	Ing 2: 10,000/ 1:100,000	Ing 2: 10,000/ 1:100,000	
	Inh 0.000003/0.0000002	Inh 0.000003/0.0000002	Inh 4: 10,000/ 9: 1,000,000	Inh 8:100,000/ 1: 100,000	
PAHs TEQ (other ^{\dagger})	Ing 0.000006	Ing 0.0000006	Ing 2:1,000,000	Ing 2:1,000,000	
	Inh 0.0000002	Inh 0.00000002	Inh <1:1,000,000	Inh <1:1,000,000	
ATSDR 1995 (Update)	Child res. ingestion dose (0.000	07) is 3,714 times less than the do	ose (2.6, Neal and Rigdon, 1967) associated with stomach	
	cancer in mice exposed to benz	o[a]pyrene ad lib in food for 30 t	to 197 days (non-cancer illnesse	es are all associated with much	
	higher doses).				
	Adult res. ingestion dose (0.000	0007) is 37,142 times less than th	ne (2.6) dose associated with sto	mach cancer in mice.	
	Inhalation res. exposure level (0.00000002) is 5,000 times less t	han the dose (0.0001, Gupta et a	al. 1993)) associated with	
	reduced lung function, abnorma	al chest x-ray, cough, bloody vor	nit, and throat and chest irritation	on, in persons exposed from 6	
	months to 6 years.				
	Child other ingestion dose (0.0	00006) is 433,333 times less than	the dose (2.6) dose associated	with stomach cancer in mice.	
	Adult other ingestion dose (0.0	000006) is 4,333,333 times less t	than the (2.6) dose associated w	ith stomach cancer in mice.	
	Inhalation other exposure level	(0.0000002) is 5,000 times less	s than the dose (0.0001, Gupta e	t al. 1993) associated with	
+	reduced lung function, abnorma	al chest x-ray, cough, bloody vor	nit, and throat and chest irritation	on, in persons exposed from 6	
Other properties include	months to 6 years.				
right-of-ways, city-	Cancer and occupational studie	s Worker exposures to high level	Is of PAHs show cancers (skin,	bladder, lung and	
owned or vacant lot	gastrointestinal) are the most si	gnificant endpoints of PAH toxic	city. Long-term worker PAH ex	posures have been linked with	
surface soil.	skin and eye irritation, photoser	nsitivity, respiratory irritation (w	ith cough and bronchitis), leuko	plakia ² , precancerous skin	
	growths enhanced by exposure	to sunlight, erythema ⁻ , skin burr	is, acheiform lesions, mild hepa	totoxicity, and haematuria.	
	Also several PAH compounds	are immunotoxic, and some supp	ress selective compounds of the	e immune system. Workers	
	dermal exposure studies indicat	te that although direct contact ma	be of concern at high exposur	re levels, they do not suggest	
	that lower levels are likely to ca	ause significant irritation (Goodf	ellow et al. 2001).		

 $[\]Delta$ Erythema nodosum is an inflammation of subcutaneous fat tissue.

[€]Hematuria is passage of blood in the urine.

 $[\]Omega$ Leukoplakia is a common, potentially pre-cancerous disease of the mouth that involves the formation of white spots on the mucous membranes of the tongue and inside of the mouth. Despite the increased risk associated with having leukoplakia, many people with this condition never get oral cancer.



Table 9. Comparison of doses calculated from highest measured subsurface soil values to lowest observable adverse effect levels(LOAELs) in animal and human medical studies. Bolded doses are above LOAEL or minimum risk level (MRL).

Chemical		Soil Doses are in mg/kg/day		theoretical increased cancer risk			
Chemiear		children's dose	adults' dose	children's	adults'		
Arsenic	(residences)	Ing 0.0007	Ing 0.00008	Ing 5:100,000	Ing 5:100,000		
		Inh 0.000003	Inh 0.000003	Inh 2:1,000,000	Inh 3:1,000,000		
Arsenic	$(other^{\dagger})$	Ing 0.0005	Ing 0.00006	Ing 3:100,000	Ing 4:100,000		
		Inh 0.000002	Inh 0.000002	Inh 2:1,000,000	Inh 3:1,000,000		
ATSDR 20	000 (Update)	Child res. ingestion dose (0.0007) is about 1/3rd the LOAEL dose (0.002, Chiou et al. 1997) associated with cerebrovascular disease and					
		cerebral infarctions. Another study showed skin effects at 0.0018 (Haque, et al., 2003). This child residential ingestion dose is also 2.3 times					
		greater than the MRL (0.0003) cal	culated from the NOAEL (0.0008) f	or adverse skin effects seen in anot	her long-term study of ingestion of		
		arsenic in drinking water. ATSDR	scientists divided this NOAEL dose	e (0.0008, Tseng et al. 1968) by 3 to	b account for human diversity in		
		calculating the MRL (0.0003). While sensitive children might experience cerebrovascular or skin effects if they had daily, long-term					
exposures to this arsenic-contaminated soil, such exposures might not occur daily since this soil is in the subsurface.					subsurface.		
		Adult res. ingestion dose is 8.75 times less than the (0.0007) dose referenced for children and 3.75 times less than the MRL. Most adults would be unlikely to experience skin or gestrointestinal bealth effects.					
		would be unlikely to experience skin of gastrointestinal nearin effects.					
1111111111011111111111111111111111111					ans inhaling arsenic Dermatitis is		
skin inflammation that may cause redness pain and occasionally itching Because the soil containing this level of arsen					s level of arsenic is not directly at		
the surface, persons are less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even less likely to have daily exposure to it; therefore, the risk of adverse health effect could be even l				could be even less likely than these			
		factors indicate.					
	Child other ingestion dose (0.0005) is 1/4 th the LOAEL dose (0.002) associated with cerebrovascular disease and cerebral infarc						
skin effects at 0.0018. This dose is 1.5 times the MRL (0.0003). As with the child residential dose, while sensitive children mi				sensitive children might experience			
gastrointestinal or skin effects if they had daily, long-term exposures to this arsenic-contaminated soil, such e			ch exposures may not occur daily				
	to this subsurface soil.						
Adult other ingestion dose is 8.75 times less than the children's other dose (0.0005) dose. We would not expect skin				expect skin or gastrointestinal			
		health effects for adults at this level of exposure.					
		Inhalation other exposure level (0.000002) is 350 times less than the amount associated with increased risk of still birth in humans (0.0007) in humans inheling grantic Dermetitie is this influence in the height has					
		and $3,500$ times less than the dose causing dermatitis (0.007) in humans inhaling arsenic. Dermatitis is skin inflammation which mig					
.	Arsenic Associated cancers: From lowest to highest dose cancer effect levels, chronic arsenic exposures in people have by						
[†] Other properties				endothelioma), urinary tract cancer	rs (bladder, kidney, ureter, and all		
include right-of-ways, urethral cancers), and intraepidermal cancers. Intraepidermal is the name for the early pre-invasive form of squamo			of squamous cell skin cancer. Pre-				
city-owned	d or vacant	invasive means that the cancer cell	ls are confined to the outermost laye	r of skin, the epidermis. At this sta	ge, the cancer cells are unlikely to		
lot surface	soil	have spread to the lymph nodes, but	ut they can spread along the skin sur	face. If left untreated, these cells c	an develop into an invasive cancer		
	55011.	and spread into the lymphatic syste	em.				



Chemical	Soil Doses are in mg/kg/day		theoretical increased cancer risk	
chemieur	children's dose	adults' dose	children's	adults'
Barium (residences)	Ing 0.007 Inh 0.00003	Ing 0.0007 Inh 0.00003	No Slope No Unit Risk	No Slope No Unit Risk
Barium (other [†])	Ing 0.03 Inh 0.0001	Ing 0.003 Inh 0.0001	No Slope No Unit Risk	No Slope No Unit Risk
ATSDR 2006 (Update) Draft	Child res. ingestion dose (0.007) is 86 times less than the MRL. Rats dosed (115 mg/kg/day, NTP 1994) with barium chloride in their food for 90 days showed increased kidney weight. The NOAEL for this study was 65 mg/kg/day. ATSDR authors used the NOAEL from this study to derive the chronic MRL of 0.6 for barium exposure by dividing the NOAEL by 100 (10 to account for extrapolation from animals to humans and 10 for human variability). Adult res. ingestion dose 0.0007) is 10 times less than the (0.007) dose referenced for children, we would not expect kidney or other health effects for most adults and children, at these exposure levels. Child other ingestion dose (0.003) is 20 times less than the MRL of 0.6. Adult other ingestion dose (0.003) is 200 times less than the MRL of 0.6. We would not expect kidney or other			
[†] Other properties include right-of-ways, city-owned or vacant lot surface soil.	residential subsurface soils. <u>Inhalation exposure level</u> Medical case reports and animal studies are inadequate to establish the heath effects of barium by inhalation. The lowest reported exposure level is 0.06 mg/m ³ /minute. Guinea pigs exposed for an unspecified amount of time to this concentration of aerosolized barium chloride solution experienced bronchoconstriction. The estimated exposure level, (0.0001) is 600 times less. <u>Barium associated cancers.</u> Barium has not been shown to cause cancer in people or animals.			



Chemical	Soil Doses are in mg/kg/day		theoretical increased cancer risk		
Chemical	children's dose	adults' dose	children's	adults'	
Dioxin TEQ (residences)	Ing 0.000000001 Inh 0.000000000004	Ing 0.0000000001 Inh 0.000000000004	Ing 8:1,000,000	Ing 8:1,000,000	
Dioxin TEQ (other ^{\dagger})	Ing 0.000000001 Inh 0.000000000006	Ing 0.0000000001 Inh 0.000000000006	Ing 9:1,000,000	Ing 1:100,000	
ATSDR 1998a (Update)	Inh 0.00000000006 Inh 0.00000000006 Child res. ingestion dose (0.00000001) is 1/120 th the dose (0.00000012, Rier et al., 1993) associated with reproductive effects (moderate endometriosis) and altered social behavior in a rhesus monkey dioxin ingestion study. This dose is also equal to the MRL of 0.00000001 calculated from the 0.00000012 dose (Schantz et al., 192) at which rhesus monkeys showed adverse neurobehavioral and developmental effects. ATSDR authors divided this dose by an uncertainty factor of 90, 3 for use of a minimal LOAEL, 3 for extrapolation from animals to humans and 10 for human variability. The results of other oral animal studies suggest that the effects that occur at the lowest levels of dioxin doses are immune and endocrine, in addition to developmental effects. People's ingestion exposures are mainly known from low levels of food contamination. Adult res. ingestion dose (0.00000001) is 10 times less than the residential children's dose. We would not expect health effects for most adults and children at these residential exposure levels. Child other ingestion dose (0.00000001) is 1/120 th the dose (0.00000012) associated with reproductive effects (moderate endometriosis) and altered social behavior in a dioxin rhesus monkey study and it too equals the MRL. The results of oral animal studies suggest that the effects that occur at the lowest levels of dioxin doses are immune, endocrine, and developmental effects. People's ingestion dose (0.000000001) is 10 times less than the residential children's dose. We would not expect health effects. People's ingestion exposures are mainly known from low levels of food contamination. Adult res. ingestion dose (0.000000001) is 10 times less than the residential children's dose. We would not expect health effects. Peopl				
right-of-ways, city-	, city- diabetes. The highest exposure levels are associated with nervous system effects, chloracne, respiratory effects, and increased risk cancer.				
owned or vacant lot surface soil.	or vacant lot Cancers Statistically significant increases in risks for all cancers were found in workers highly exposed to dioxins with longer la periods. Although the estimated Standardized Mortality Ratios [†] are low [†] , they are consistent across studies with the highest dioxin exposures. The evidence linking peoples' doses with site-specific cancers is weaker, with some data suggesting a possible relationship between soft-tissue sarcoma, non-Hodgkin's lymphoma, or respiratory cancers including lung cancer. Animal studi have also shown associations with liver thyroid and skin cancer.				

[†] Standardized Mortality / Morbidity Ratio (SMR) is a widely used method of reporting death or disease which adjusts for differences in age and sex across regions. It is a measure of premature mortality. Instead of giving an adjusted rate, the SMR gives a ratio that is a direct comparison with a standard (e.g. the entire state).



Chamical	Soil Doses are in mg/kg/day		theoretical increased cancer risk		
Chennear	children's dose	adults' dose	children's	adults'	
PAHs TEQ (residences)	Ing 0.0002/0.00004 Inh 0.000001/0.00000007	Ing 0.00003/0.000002 Inh 0.000001/0.00000007	Ing 8:100,000/ 5: 1,000,000 Inh 1:10,000/ <1:1,000,000	Ing 8:100,000/ 6: 1,000,000 Inh 9:100,000/ 4:1,000,000	
PAHs TEQ (other [*])	Ing 0.000002	Ing 0.0000002	Ing <1:1,000,000	Ing <1:1,000,000	
	Inn 0.00000008	Inn 0.00000008	Inh <1:1,000,000	Inn 2:1,000,000	
ATSDR 1995 (Update)	ATSDR 1995 (Update) We would not expect non-cancer health effects for most adults and children at these residential or other exposure <u>Child res. ingestion dose</u> (0.0002) is 13,000 times less than the dose (2.6, Neal and Rigdon, 1967) associated with cancer in mice exposed to benzo[a]pyrene ad lib in food for 30 to 197 day (non-cancer illnesses are all associated bicker down)				
	higher doses).Adult res. ingestion dose (0.00003) is 6.6 times less than the residential children's dose.Inhalation res. exposure level (0.000001) is 100 times less than the dose (0.0001, Gupta et al. 1993) associated with reducedlung function, abnormal chest x-ray, cough, bloody vomit, and throat and chest irritation, in persons exposed from 6 months to6 years.Child other ingestion dose (0.00002) is 1,300,000 times less than the dose (2.6) dose associated with stomach cancer in mice.Adult other ingestion dose (0.000002) is 10 times less than the residential children's dose.				
*Other properties include right-of-ways, city- owned or vacant lot surface soil.	 <u>Inhalation other exposure level</u> (0.00000008) is 12,500 times less than the lost (0.0001) associated with reduced lung function, abnormal chest x-ray, cough, bloody vomit, and throat and chest irritation, in persons exposed from 6 months to 6 years. <u>Cancer and occupational studies</u> Worker exposures to high levels of PAHs show cancers (skin, bladder, lung and gastrointestinal) are the most significant endpoint of PAH toxicity. Long-term worker PAH exposures have been linked with skin and eye irritation, photosensitivity, respiratory irritation (with cough and bronchitis), leukoplakia^Ω, precancerous skin growths enhanced by exposure to sunlight, erythema^Δ, skin burns, acneiform lesions, mild hepatotoxicity, and haematuria[€]. Also several PAH compounds are immunotoxic, and some suppress selective compounds of the immune system. Workers' dermal exposure studies indicate that although direct contact may be of concern at high exposure levels, they do not suggest that lower levels are likely to cause significant irritation (Goodfellow et al. 2001). 				

 $[\]Omega$ Leukoplakia is a common, potentially pre-cancerous disease of the mouth that involves the formation of white spots on the mucous membranes of the tongue and inside of the mouth. Despite the increased risk associated with having leukoplakia, many people with this condition never get oral cancer.

 $[\]Delta$ Erythema nodosum is an inflammation of subcutaneous fat tissue.

[€]Hematuria is passage of blood in the urine.



Appendix C—Safe Gardening Card





Lincoln Park Complex, Durrs Neighborhood Offsite Soil Samples Health Consultation



Certification

The Florida Department of Health, Bureau of Community Environmental Health prepared this Health Consultation under a cooperative agreement with the Agency for Toxic Substances and Disease Registry. Florida DOH followed approved methodologies and procedures existing at the time the health consultation was begun. The Cooperative Agreement Partner completed editorial review.

Technical Project Officer CAT SPAB, DHAC, ATSDR

The Division of Health Assessment and Consultation, ATSDR, reviewed this health consultation, and concurs with its findings.

Alan Yarbrough Team Lead, CAT, SPAB, DHAC, ATSDR